

Parallel between Offline-SPE-LC-MS and Direct Injection LC-MS Methods for Acrylamide Detection in Drinking Water at Parts Per Trillion Level

FLORENTINA LAURA CHIRIAC, IULIANA PAUN, FLORINELA PIRVU, LILIANA CRUCERU, LUOANA FLORENTINA PASCU, TOMA GALAON*

National Research and Development Institute for Industrial Ecology ECOIND, 71-73 Drumul Podu Dambovitei Str., 060652, Bucharest, Romania

The presence of acrylamide, a potential human carcinogen, in treated drinking water has been a long-time concern for public health agencies. Due to this, acrylamide is included among the substances to be monitored in drinking water with a stated maximum admissible level of 0.1 µg/L in Europe. The necessity for ultra-trace level detection of acrylamide has led to development of an Offline-SPE-LC-MS method that involves SPE extraction of acrylamide using a strong adsorbent like activated charcoal followed by concentration. Simultaneously, a simple and rapid large volume Direct Injection LC-MS method was optimized from the previous method to detect acrylamide without the resource and time-consuming SPE step. A comprehensive comparison between the efficiency and sensitivity of the two methods is made. Optimization of LC-MS parameters (column temperature, mobile phase composition and flow, collision energy, fragmentor voltage, capillary voltage or MRM transitions) allowed high detection sensitivity. Due to Acrylamide high polarity, a Synergi Fusion-RP column with amide polar embedded groups was chosen to achieve reasonable retention and peak shape for acrylamide with a short run-time of only 3 min. The two methods provided excellent sensitivity, with reasonably low LOQ values of 2 ng/L for the Offline-SPE-LC-MS method and 30 ng/L for the Direct Injection LC-MS method. Intra-day and inter-day precision (RSD%) were 7.3 and 10.3%, for the Offline-SPE-LC-MS method and 2.4% and 6.2%, respectively for the Direct Injection LC-MS method. The analysis of acrylamide was successfully performed in drinking water from different location consumers and water treatment plants in Romania. Acrylamide was found in drinking water samples with levels between 5.2 and 35.7 ng/L using the Offline-SPE-LC-MS method. The direct injection method can only be used for acrylamide levels higher than its LOQ value (30 ng/L).

Keywords: acrylamide, offline-SPE-LC-MS, direct injection LC-MS, drinking water

Polyacrylamide is widely used in water treatment plants for the flocculation of suspended material and in many other products and activities: cosmetics [1], pesticides [2], grout, cement [3], explosives, food manufacturing [4], crude oil production [5], coatings for home appliances, building materials, adhesives [6], automotive parts and printing inks. The main source of acrylamide (AA) in the environment is the release of residual monomer from polyacrylamide used as flocculant in the treatment of raw water. Another significant source of acrylamide is the thermally processed food and it can be formed in two ways: (i) by amino acids interaction (asparagine) with sugars in the presence of heat and (ii) food fats oxidation during frying process and their conversion to acrylic acid and acrolein. These molecules will interact with asparagine, in the presence of heat and form unusual amounts of acrylamide [7,8].

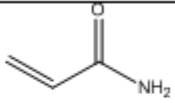
Acrylamide is considered as possible carcinogenic to humans and classified as a Group 2A carcinogen by the International Agency for Research on Cancer (IARC) [9] and the World Health Organization (WHO) guideline value associated with a lifetime cancer risk is 0.5 µg/L in drinking water [10]. Concerning its use in drinking water supplies, the US EPA set a maximum contaminant level goal (MCL) for acrylamide of zero and requires water suppliers to demonstrate that the AA monomer is present at concentrations lower than 0.5 µg/L [11]. In Europe, the REACH regulation limited the acrylamide content in commercialized products to < 0.1% for grouting applications after 2012 [12]. Acrylamide has also been

regulated in European countries by the Drinking Water Directive which imposes a maximum admissible concentration of 0.1 µg/L for water intended for human consumption [13].

The very low concentration of different types of pollutants in environmental samples, usually at parts per trillion levels, requires the isolation of interest analytes by extraction procedures and extremely sensitive analytical methods [14-16]. Because of its physico-chemical properties, such as low molecular weight (M.W. = 71.1 g/mol) and significant polarity (table 1), this molecule is difficult to separate and to detect using either GC-MS or LC-MS methods.

Acrylamide has been determined in environmental water samples by gas chromatography using electron capture (GC-ECD) or mass spectrometric (GC-MS) detection [17-20]. These methods require derivatization and liquid-liquid

Table 1
ACRYLAMIDE PHYSICO-CHEMICAL PROPERTIES AND STRUCTURE

Property	Acrylamide
Chemical Structure	
Molecular Mass (g/mol)	71.08
log Kow	-0.67
Density (g/cm ³)	1.13
Melting point (°C)	84.5
Water Solubility (%)	> 40

* email: tomagalaon@yahoo.com

extraction before analysis. However, the efficiency of derivatization is often variable and the process is time consuming and requires usually additional LLE extraction before GC separation. To avoid the derivatization procedure, high-performance liquid chromatography (HPLC) is a suitable method for determination of acrylamide. While HPLC with UV detection suffers from poor detection limits (3-10 µg/L) [21,22] due to lack of chromophores, liquid chromatography tandem mass spectrometry (LC-MS/MS) was more recently employed due to its high selectivity and increased sensitivity. The trace amounts of acrylamide present in drinking water samples, involves the necessity to concentrate the samples before LC-MS analysis, which might also be problematic. Due to its high-water solubility (40%), AA is poorly retained on most solid-phase extraction (SPE) adsorbents [23].

The objective of the present study was to develop and validate two methods - an Offline SPE-LC-MS and a Direct Injection LC-MS method, both able to detect low levels of acrylamide in drinking water, and to compare their efficiency and sensitivity (LOQ).

Experimental part

Reagents and chemicals

Standards of Acrylamide (AA) and isotopically labeled Acrylamide ($^{13}\text{C}_3$ -Acrylamide, $^{13}\text{C}_3$ -AA) which was used as internal standard were purchased from Sigma-Aldrich. Methanol and acetonitrile of HPLC-grade purity were acquired from Merck and Formic Acid of 99% purity from Supelco. SPE activated charcoal cartridges (Supelclean Coconut Charcoal - 2 g / 6 mL) were acquired also from Supelco (Bellefonte, PA, USA).

LC-MS instrumentation and conditions

Analyses were performed using an Agilent 1260 series LC system (Waldbronn, Germany) coupled with an Agilent 6410B triple-quadrupole mass spectrometer with electrospray ionization source (ESI). The chromatographic column was a Synergi Fusion RP (150 x 2.0 mm, 4.0 µm) from Phenomenex, the stationary phase being a C18 chain modified with polar embedded amide groups. All experiments were performed in isocratic elution conditions at a flow-rate of 0.2 mL/min. Mobile phase consisted of a binary mixture of 30% Aq. Formic acid (0.1%) and 70% MeOH. Optimized injection volumes of 10 µL for the Offline SPE-LC-MS method and 50 µL for the Direct injection LC-MS method were used with water/MeOH (1:1) as sample diluent. MS detection was achieved using Multiple Reaction Monitoring (MRM) acquisition mode. Full-Scan MS spectra were acquired in the range 20 - 100 Da to establish MRM transitions for AA and internal standard. Retention time, MRM transitions, fragmentor voltages, collision energies and other MS parameters are given in table 2. ESI ionization source was operated in positive mode with 300°C as the drying gas temperature, 6 L/min drying gas flow, 40 psi nebulizer pressure and 6000 V capillary voltage. SPE extraction of acrylamide and internal standard from drinking water was done using the semi-Automated extractor Dionex AutoTrace 280 from Thermo Scientific.

Results and discussions

LC separation optimization

Considering the significant polarity of the investigated analyte (log Kow = -0.67) and its low molecular weight, it was expected that this compound would not have significant retention on hydrophobic C18 columns and also might have problems of sensitivity for MS detection. For this reason, a C18 polar embedded column with amide polar groups was chosen to obtain a reasonable retention time and peak symmetry for the acrylamide and IS peaks. The C18 ligand gives normal retention for hydrophobic compounds while the polar embedded group provides enhanced retention for highly polar analytes like acrylamide (log Kow = -0.67). Various mobile phase compositions were tested for elution using different organic solvents (ACN, MeOH) and also different proportions between the aqueous and the organic solvent. The chosen mobile phase composition Aq. 0.1% formic acid / MeOH = 30 / 70 (v/v) generated good peak shape and provided short retention time for AA peak (< 3 min). Other LC instrumental parameters taken into account to increase AA and $^{13}\text{C}_3$ -AA peak area, peak efficiency and symmetry (peak shape) were the injection volume and column temperature. Injection volume range between 5 and 75 µL and column temperature range from 15 to 45°C were tested. A reasonable peak shape for AA and $^{13}\text{C}_3$ -AA was achieved for injection volumes up to 50 µL. Above this value peak distortion appeared for both AA and $^{13}\text{C}_3$ -AA. It is important to be mentioned that an injection volume of 50 µL is very high for the used column (2.0 mm i.d.). Injection of large volumes in RP-LC can be employed if sample diluent is weaker than mobile phase [24]. Peak symmetry and peak efficiency of AA and $^{13}\text{C}_3$ -AA increased with temperature increase from 15 to 45°C and the latter value was thus chosen. Optimized conditions of the LC parameters allowed a short run-time of only 3 min. Figure 1 shows the MS/MS extracted chromatograms of Acrylamide and $^{13}\text{C}_3$ -Acrylamide standard mixtures obtained using the two methods.

MS detection optimization

Method development for acrylamide detection implied optimization of all MS parameters to obtain lowest possible quantitation limit (LOQ) which is necessary to determine extremely low levels of acrylamide expected to occur in drinking water (ng/L). Thus, all detection parameters of the triple quadrupole MS detector (QQQ) were modified to obtain maximum sensitivity (fig. 2). All electrospray ionization source parameters were also modified in a wide range to obtain highest ionization efficiency for the analytes in terms of peak area (detector response) but also in terms of signal-to-noise ratio (S/N). Collision energy (CE) applied in the collision cell (Q2) to the precursor ion to generate the product ion of the MRM transition was varied in the range 5-15 V. The highest peak area value was obtained for a collision energy of 10V. Using the optimum collision energy values, the MS optimization was continued with fragmentor voltage (voltage used to accelerate, focus and transfer ions from the ionization source to the first

Compound	MRM transitions	Fragmentor voltage (V)	Collision energy (V)	Cell Voltage (V)	Polarity
$^{13}\text{C}_3$ -Acrylamide	75 → 58 (Q)	80	10	1	+
Acrylamide	72 → 55 (Q)	80	10	1	+
	72 → 54 (q)	80	10	1	+

Table 2
MS PARAMETERS FOR THE DETECTION OF ACRYLAMIDE AND LABELED INTERNAL STANDARD (MRM TRANSITIONS, FRAGMENTOR VOLTAGE, COLLISION ENERGY AND OTHER MS PARAMETERS)

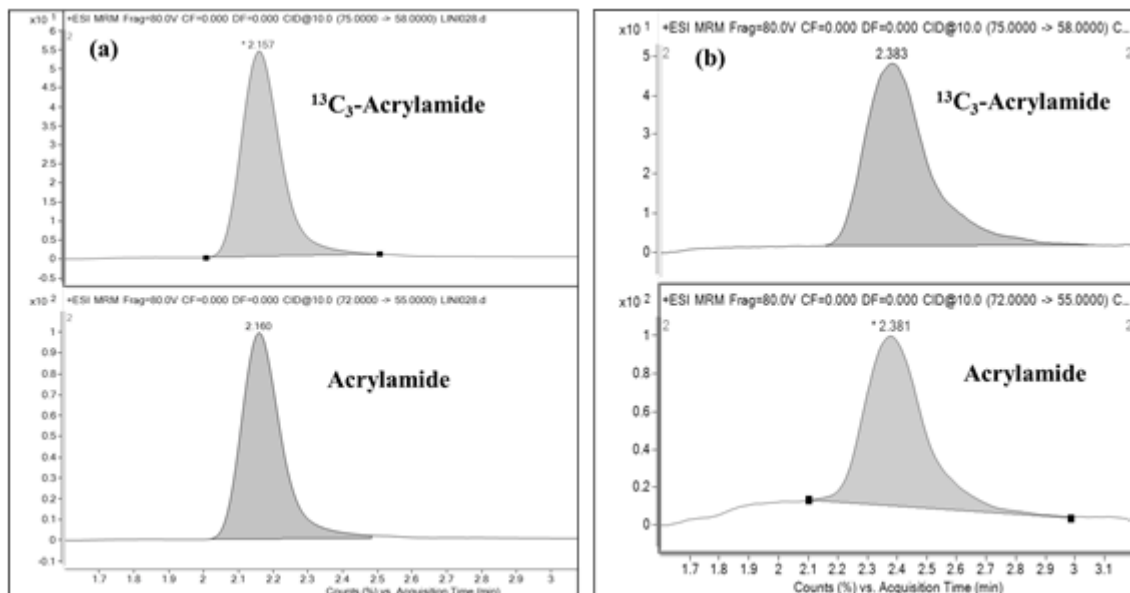


Fig. 1. MRM overlaid extracted ion chromatograms of (a) 50 µg/L AA and 100 µg/L $^{13}\text{C}_3$ -Acrylamide (SI) using the Offline-SPE-LC-MS method and (b) 0.5 µg/L AA and 1 µg/L $^{13}\text{C}_3$ -Acrylamide (SI) using the Direct injection LC-MS method

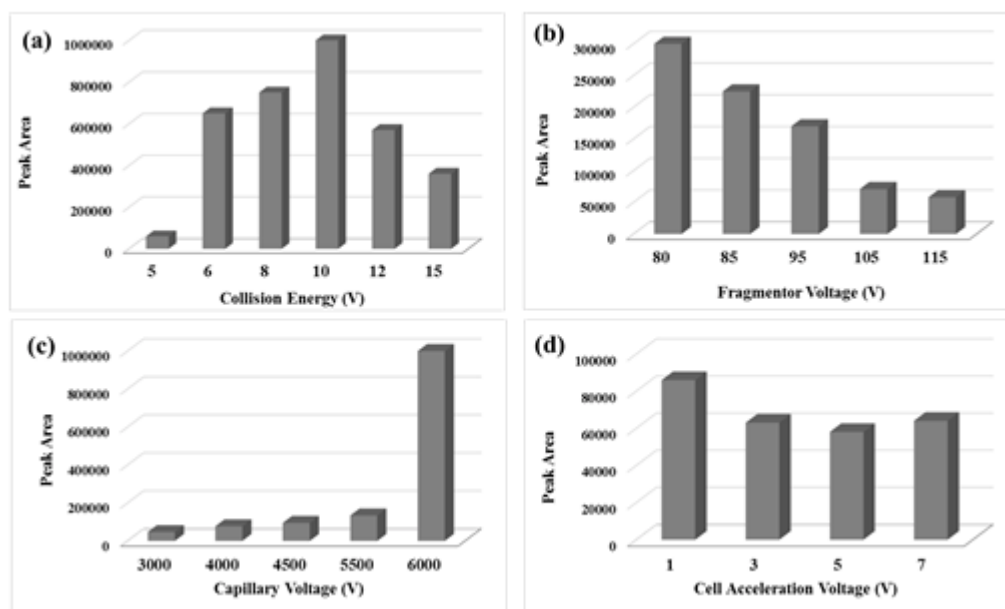


Fig. 2. Peak area variation of Acrylamide during MS optimization function of: (a) Collision energy; (b) Fragmentor voltage; (c) Capillary Voltage and (d) Cell accelerator voltage

quadrupole). The values selected for testing were between 80 and 115 V (5 V steps). Fragmentor voltage of 80 V generates the highest transfer of acrylamide precursor ion to the first quadrupole. The same procedure was applied for the capillary voltage in the range 3000 - 6000 V and to the cell accelerator voltage in the range 1 - 7 V. The final chosen values (6000 V capillary voltage and 1 V for cell accelerator voltage) generate maximum peak area as can be seen in figure 2.

Following MS detection optimization procedure, the parameters which generated maximum sensitivity were selected and are given in table 3.

Offline-SPE extraction

The optimization of the LC-MS analysis method was followed by optimization of the solid phase extraction parameters (using semi-automated SPE extractor) to allow concentration of acrylamide from drinking water samples at ppt level up to ppb level which represents the usual instrumental quantitation limit for MS/MS detection with Triple Quadrupole analyzers. Acrylamide was extracted from 250 mL water samples using activated charcoal SPE cartridges. This adsorbent is usually efficient for highly polar molecules like AA and $^{13}\text{C}_3$ -AA. The cartridges were preconditioned with 5 mL of methanol and 5 mL of

MS parameters	Optimized values
Ionization source:	Electrospray (positive mode) - ESI(+)
Drying gas temperature:	300°C
Drying gas flow:	6 L/min
Nebulizer pressure:	40 psi
Capillary Voltage:	6000 V
Collision Energy:	10 V
Fragmentor voltage:	80 V
MS mode:	Multiple Reaction Monitoring (MRM)
Dwell time:	160 msec
Cell Accelerator Voltage:	1 V

Table 3
OPTIMIZED MS PARAMETERS THAT GENERATE
MAXIMUM SENSITIVITY FOR ACRYLAMIDE

ultrapure water at 5 mL/min. Before extraction, 1 mL of $^{13}\text{C}_3$ -AA (100 µg/L) was added as surrogate IS to each water sample. The samples were passed through the SPE cartridges. Before elution, the cartridges were dried for 30 min. Elution was done with 10 mL of methanol at a flow rate of 1 mL/min. Under a gentle nitrogen stream, the extracts were evaporated to dryness and re-dissolved with 1.0 mL of water/MeOH (1:1) and transferred into injection vials prior to LC-MS analysis. Thus, a concentration factor of 250 was obtained using the SPE extraction.

Method validation

To account for their performance, the developed Offline-SPE-LC-MS and Direct injection LC-MS methods were validated with respect to specificity, linearity, precision, accuracy and limit of quantitation. MS detector response was tested and proved linear in the range 1 - 200 µg/L for Offline-SPE-LC-MS method and 0.05 - 2.0 µg/L for the direct injection LC-MS method. Both methods generated high correlation coefficients ($R^2 > 0.999$) (fig. 3).

Instrumental LOQs for both methods were determined by injecting decreasing concentrations of acrylamide solutions until a S/N of 10 was obtained (table 5). Intra-day and inter-day method precision was tested on 6 replicate samples by spiking 50 ng/L for Offline-SPE-LC-MS and 0.1 µg/L acrylamide for the Direct injection method. Results obtained after validation procedure are given in Table 4. The methods proved to be precise with RSD% values of 7.3% (intra-day) and 10.3% (inter-day) for Offline-SPE-LC-MS method and 2.4% (intra-day) and 6.2% (inter-day) for Direct injection LC-MS method.

Analyte recovery was calculated with internal standard correction as can be observed in table 5. Overall method LOQs were determined as 2 ng/L for Offline-SPE-LC-MS and 30 ng/L for Direct injection LC-MS method.

Table 4

RESULTS OBTAINED FOR INTRA-DAY AND INTER-DAY PRECISION BY SPIKING 50 ng/L ACRYLAMIDE FOR OFFLINE-SPE-LC-MS AND 0.1 µg/L FOR THE DIRECT INJECTION LC-MS METHOD

Offline-SPE-LC-MS RSD%		Direct injection LC-MS RSD%	
Intra-day	Inter-day	Intra-day	Inter-day
7.3	10.3	2.4	6.2

Comparing the performance of the two methods it can be observed that although the Offline-SPE-LC-MS one allows the detection of ultra-trace levels of acrylamide (LOQ 2 ng/L), it requires a time-consuming SPE extraction procedure which takes about 3 hours being overall more expensive. The Direct injection LC-MS method on the other hand is more rapid (no sample preparation step), cost-efficient but allows quantitation of AA only for concentration levels higher than its LOQ value of 30 ng/L. Nonetheless this value is more than 3 times lower than the European maximum admissible concentration (0.1 µg/L). The Direct injection method is more precise than the Offline-SPE method due to lack of additional preparation steps and does not suffer from analyte loss during extraction as opposed to the latter (40.2% absolute recovery). Acrylamide loss during SPE is compensated by the labeled internal standard addition, but the loss affects method LOQ.

Acrylamide occurrence in drinking water

The developed Offline-SPE-LC-MS method was used for detection of acrylamide ultra-trace levels in drinking water samples collected from ten different private consumers (S1-S10) and five water treatment plants (S11-S15) all located in Neamt County, Romania. The values determined for acrylamide by both Offline-SPE-LC-MS and Direct injection LC-MS methods are shown in table 6.

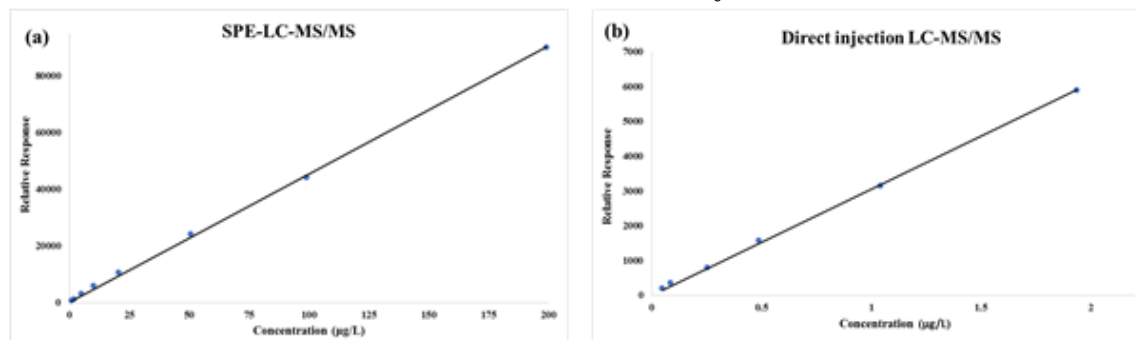


Fig. 3. Linear MS detector response in the range (a) 1 - 200 µg/L for Offline-SPE-LC-MS method and (b) 0.05 - 2.0 µg/L for Direct injection LC-MS method

Offline-SPE-LC-MS					Direct injection LC-MS	
IQL (µg/L)	LOD (ng/L)	LOQ (ng/L)	Absolute recovery (%)	Recovery with IS (%)	LOD (ng/L)	IQL = LOQ (ng/L)
0.4	0.6	2.0	40.2	98.4	9	30

Samples	Offline-SPE-LC-MS (ng/L)	Direct injection LC-MS (ng/L)
S1	17.7	< LOQ
S2	15.9	< LOQ
S3	18.6	< LOQ
S4	15.2	< LOQ
S5	19.4	< LOQ
S6	7.9	ND
S7	5.2	ND
S8	5.5	ND
S9	< LOQ	ND
S10	35.7	38.6
S11	3.8	ND
S12	4.2	ND
S13	8.7	ND
S14	11.4	< LOQ
S15	34.7	36.9

Table 5
ANALYTE RECOVERY, INSTRUMENTAL QUANTITATION LIMITS (IQLs), DETECTION AND QUANTITATION LIMITS FOR THE TWO METHODS (LODs AND LOQs)

Table 6
ACRYLAMIDE CONCENTRATION (ng/L) DETECTED IN DRINKING WATER SAMPLES

Using the Offline-SPE-LC-MS method, Acrylamide was determined to be present in almost all analyzed samples (except for S9) with concentration values between 5.2 and 35.7 ng/L. Due to low levels of acrylamide in tested drinking water samples, the Direct injection LC-MS method was able to quantitate acrylamide only in S10 and S15 samples because of the higher LOQ value (30 ng/L). All analyzed samples (private consumers and water treatment plants) generated AA levels below the maximum admissible concentration of 0.1 µg/L. Obtained values from Romania drinking water samples, are of the same order of magnitude with those obtained in other studies from Europe [25,26]

Conclusions

In this study two LC-MS methods were developed and validated for detection of acrylamide in drinking water samples. The first method was an Offline-SPE-LC-MS method (LOQ 2 ng/L) and the second was a Direct Injection LC-MS method (LOQ 30 ng/L) derived from the former but without the SPE extraction and with an enlarged injection volume (50 µL). The Offline-SPE-LC-MS method is more sensitive than the Direct injection LC-MS method but requires SPE extraction and concentration which means additional costs and time per analysis. LC-MS analysis was done in less than 3 min for both methods. Chromatographic conditions and MS parameters were optimized in order to obtain highest method sensitivity. In drinking water samples, the recovery of the Offline-SPE-LC-MS method was 98% due to surrogate internal standard correction. For the solid phase extraction method, the values obtained for intra-day precision was 7.3% and 10.3% for intra-day precision respectively. Regarding the direct injection method precision, RSD values were 2.4% for inter-day and 6.2% for intra-day precision. Analysis of drinking water samples was done using both methods. Acrylamide was found in drinking water samples with levels between 5.2 and 35.7 ng/L using the more sensitive Offline-SPE-LC-MS method. Direct injection method can be used for Acrylamide levels higher than its LOQ (0.03 µg/L) which is 3 times lower than the maximum admissible concentration stated in Europe (0.1 µg/L).

Acknowledgements. The authors acknowledge the financial support offered by The National Research Program Nucleu through Agreement no. 38N/2018, Project code PN 18 05 01 01.

References

- CALO, E., KHUTORYANSKIY, V. V., Eur. Polym. J., **65**, 2015, p. 252.
- SOJKA, R. E., BJORNEBERG, D. L., ENTRY, J. A., LENTZ, R. D., ORTS, W. J., Adv. Agron., **92**, 2007, p. 75.
- RAI, U. S., SINGH, R. K., Mat. Sci. Eng. A., **392**, 2005, p. 42.
- NEGOITA, M., ADASCALULUI, A., IORGA, E., CATANA, L., CATANA, M., BELC, N., Rev. Chim. (Bucharest), **66**, no. 4, 2015, p. 464.
- BAI, B., ZHOU, J., YIN, M., Petrol. Explor. Develop., **42**, no. 4, 2015, p. 525.
- BAIT, N., GRASSL, B., BENABOURA, A., DERAİL, C., J. Adhes. Sci. Technol., **27**, no. 9, 2013, p. 1032.
- ROMMENS, C. M., YAN, H., SWORDS, K., RICHAEI, C., YE, J., Plant. Biotechnol. J., **6**, no. 8, 2008, p. 843.
- S. E. K. TEKKELI, C. ONAL, A. ONAL, Food Anal. Methods, **5**, 2012, p. 29.
- *** International Agency for Research on Cancer. Some industrial chemicals. Lyon: IARC, **60**, 1994, p. 389.
- *** Guidelines for Drinking Water Quality, Recommendations, WHO, Geneva, **1**, 1993, p. 72.
- *** US Environmental Protection Agency, National primary drinking water regulation. Washington DC, 2002.
- *** European Parliament, European Council Registration, Evaluation, Authorisation and Restriction of Chemicals (REACH), establishing a European Chemicals Agency. no 1907, 2006.
- *** European Council, Directive 98/83/EC of 3 November 1998 on the quality of water intended for human consumption. Off. J. Eur. Commun. L330. European Union Brussels, 1998.
- PAUN, I., IANCU V. I., CRUCERU, L., NICULESCU, M., CHIRIAC, F. L., Rev. Chim.(Bucharest), **69**, no. 1, 2018, p. 27.
- CHIRIAC, F. L., CRUCERU, L., NICULESCU M., PASCU, L. F., LEHR, C. B., GALAON, T., REV.CHIM. (Bucharest), **68**, no. 8, 2017, p. 1685.
- MEDVEDOVICI, A., ALBU, F., SORA, I.D., UDRESCU, S., GALAON, T., DAVID, V., Biomed. Chromatogr., **23**, 2009, p. 1092.
- ZHANG, Y., ZHANG, G. Y., ZHANG, Y. J. Chromatogr. A, **1075**, 2005, p. 1.
- ***Method 8032A: Acrylamide by gas chromatography, U.S. EPA: Washington, DC, 1996.
- YAMINI, Y., GHAMBARIAN, M., ESRAFI, A., YAZDANFAR, N., MORADI, M., Int. J. Environ. Anal. Chem., **92**, 2012, p. 1493.
- *** Method 603: Acrolein and acrylonitrile, U.S. EPA, 1984.
- WEIDEBORG, M., KALLQVIST, T., ODEGARD, K. E., SVERDRUP, L. E., VIK, E. A., Water Res., **35**, 2001, p. 2645.
- SORA, I., GALAON, T., UDRESCU, S., NEGRU, J., DAVID, V., MEDVEDOVICI, A., J. Pharm. Biomed. Anal., **43**, 2007, p. 1437
- ROSEN, J., NYMAN, A., HELLENAS, K. E., J. Chromatogr. A, **1172**, 2007, p. 19
- DAVID, V., GALAON, T., ABOUL-ENEIN, H.Y., J. Chromatogr. A, **1323**, 2014, p. 115.
- DEARMOND, P. D., DIGOREGORIO, A., L., Anal. Bioanal. Chem., **405**, 2013, p. 4159.
- W. J. BACKE, V. YINGLING, T. JOHNSON, J. Chromatogr. A, **1334**, 2014, p.72.

Manuscript received: 17.08.2018